

### **Ozone impacts on biodiversity**

Felicity Hayes and colleagues UK Centre for Ecology & Hydrology

fhay@ceh.ac.uk



### **Overview**

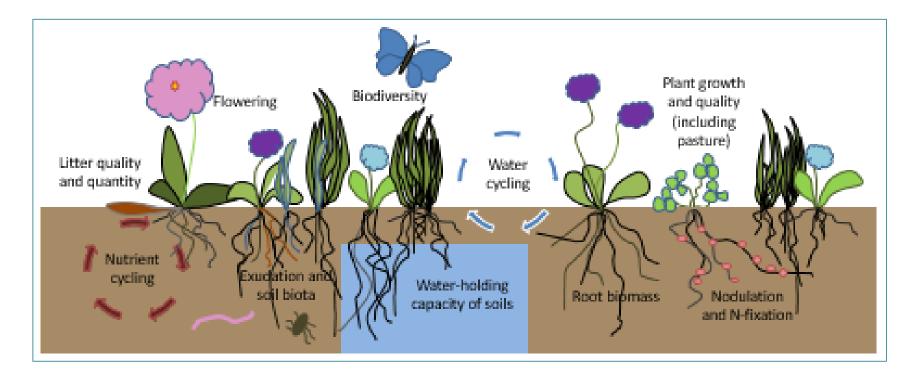
Impacts on plants

Direct impacts Indirect impacts

Impacts on insects

Direct impacts Indirect impacts

Impacts on other species





### **Ozone impacts on growth of individual species**

Number of species with ozone effects on growth						
Plant group	Reduction	Stimulation	No effect			
Forbs	85 (68)*	13 (11)	79			
Grasses	27 (20)	6 (3)	42			
(Bi)annuals	31 (23)	3 (2)	21			
Perennials	75 (60)	16 (12)	103			
Trees	70 (55)	2 (0)	37			
Deciduous	40 (32)	2 (0)	19			
Evergreen	34 (28)	0	23			
Conifers	19 (16)	0	17			
Broadleaved	56 (45)	2 (0)	25			

\* Values within brackets indicate a response of more than 15%.

Bergmann et al. (2015):

- Forbs and deciduous trees tend to be more responsive to ozone than grasses and coniferous trees.
- Although several ozone-sensitive plant families were identified, in general no significant relationship between plant traits and ozonesensitivity was found. However, in some individual studies included in the review, such relationships have been reported (Hayes et al., 2007; Jones et al., 2007)



### **Reduced plant growth and increased senescence**



Although *Dactylis glomerata* showed some response to ozone for shoots, there had been a large reduction in below-ground allocation

Current Background

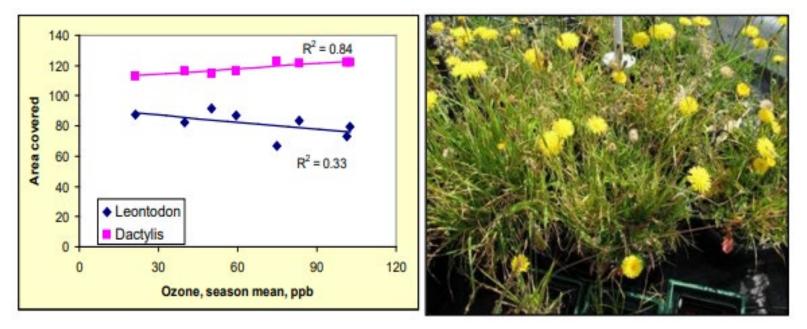
Well-watered







### Reduced plant growth = altered species composition



Some species are more sensitive than others.

Changes in species composition can occur over time

(in this example, more grass and fewer forbs with flowers)



### Risk of ozone impacts to biodiversity (based on responses of individual species)

EUNIS habitat	Abbreviated name	Mean no. of spp. in habitat	Mean % of species tested for O <sub>3</sub> sensitivity	No. of O <sub>3</sub> - responsive species*	% of O <sub>3</sub> - responsive species*
D2	Valley and transition mires	25.0	20.0	3.0	60.0
I1	Arable and market gardens	57.3	23.8	7.5	56.4
I2 E5	Cultivated gardens, parks Woodland fringes	31.0 58.9	29.0 22.8	5.0 6.8	55.6 51.1
E3 E4	(Sub)alpine grasslands	70.0	21.4	7.0	46.7
E3	Seasonally wet grasslands	63.0	25.1	6.6	43.4
B1	Coastal dunes, sandy shores	49.0	25.5	5.0	42.8
E1	Dry grasslands	86.9	26.5	10.1	41.9
E2	Mesic grasslands	71.7	35.9	10.3	41.4
B3	Rock cliffs and shores	47.3	25.7	5.0	40.1
F3	Montane scrub	61.5	23.9	6.0	38.7
D5	Sedge and reed beds	40.0	22.5	3.0	33.3
F4	Temperate shrub heathland	65.0	22.1	4.0	27.2

Using data for individual species exposed to ozone in experimental conditions.

Following approach Mills et al., 2007, EUNIS habitats most sensitive to ozone were identified. However, now based on response individual species to 24hr mean ozone.

#### Drawback: Northern European bias

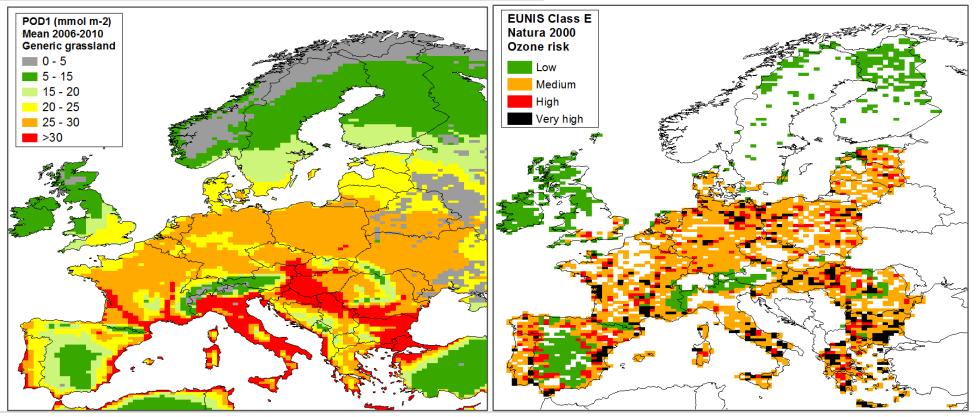
\* Species for which above-ground biomass was either reduced or stimulated.



### Risk of ozone impacts to biodiversity (based on abundance and exposure)

Grassland area in gric	Grassland area in grid	Mires, bogs, fen area in	POD <sub>1</sub> grass (mmol m <sup>-2</sup> )*	<5	5 - 15	15 - 20	20 - 25	25 - 30	>30
	cell (%)	grid cell (%)	RISK	1	2	3	4	5	6
	0.5 – 5	<1	1	1	2	3	4	5	6
	5 - 10	1-5	2	2	4	6	8	10	12
	>10	>5	3	3	6	9	12	15	18

Natura 2000 grassland at highest risk from ozone, where phytotoxic ozone doses are medium to high and habitat area is high.



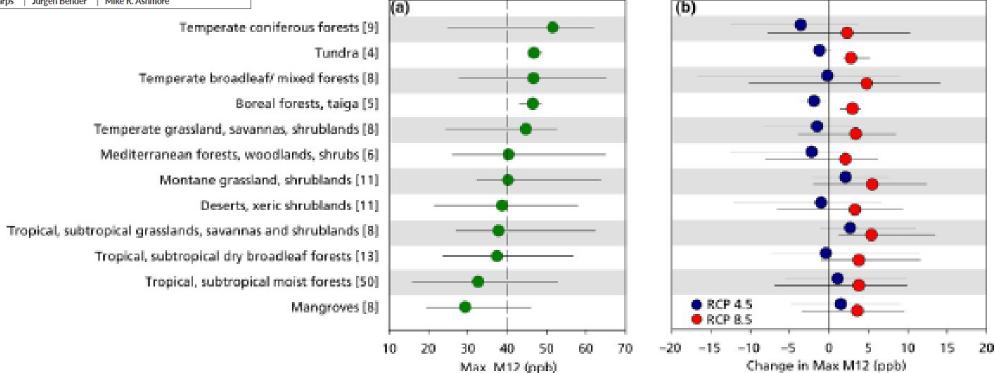


### Global risk of ozone impacts to biodiversity (based on exposure)

REVIEW

Current and future ozone risks to global terrestrial biodiversity and ecosystem processes

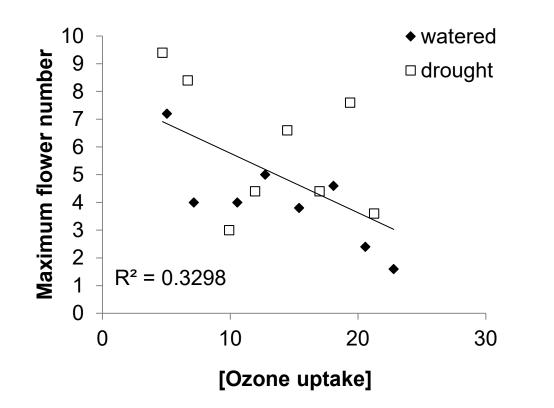
Jürg Fuhrer<sup>1</sup> | Maria Val Martin<sup>2</sup> | Gina Mills<sup>3</sup> | Colette L. Heald<sup>4</sup> | Harry Harmens<sup>3</sup> | Felicity Hayes<sup>3</sup> | Katrina Sharps<sup>3</sup> | Jürgen Bender<sup>5</sup> | Mike R. Ashmore<sup>6</sup>



Simulated  $O_3$  exposure (Max M12) in 2000 grouped by biome (left) and changes in simulated ozone exposure (2000-2050) under RCP4.5 and RCP 8.5 (right)

UK Centre for Ecology & Hydrology

### **Reduced flower numbers**





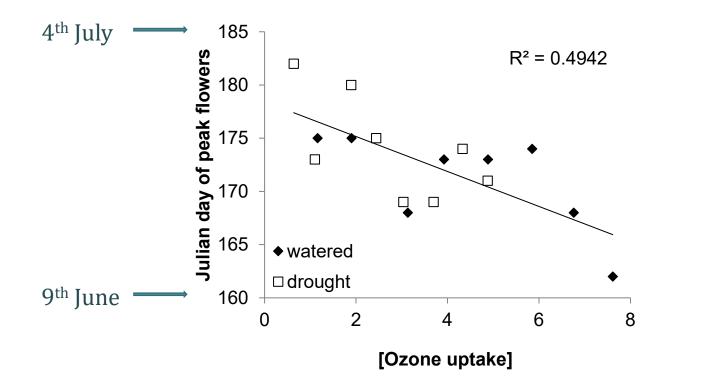
#### Harebells in grassland communities

- The number of flowers was reduced with increased ozone
- Reduced watering did not protect plants from ozone

Hayes et al. 2012. Environmental Pollution 16:40-47



### **Changes in timing of flowering**





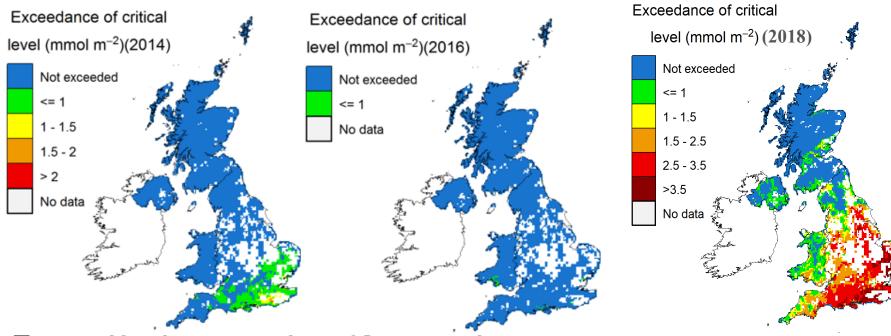
### Lotus corniculatus in grassland communities

- The date of maximum flowering was increasingly earlier with increased ozone
- Reduced watering did not protect plants from ozone
- Ecological consequences for bees/butterflies

Hayes et al. 2012. Environmental Pollution 16:40-47



### **Flower number - grasslands**



□ Critical level set at 10% loss of flower number

	Country	2014		20	16	2018	
		Area ex. (%)	Loss (%)	Area ex. (%)	Loss (%)	Area ex. (%)	Loss (%)
	England	16.3	8.3	0.4	7.5	74.8	12.8
	Wales	0	6.5	0.6	7.4	62.9	10.6
	Scotland	0	3.2	0	4.0	4.5	6.9
UK Centre	NI	0	4.9	0	7.1	18.2	9.3
Ecology 8	UK	3.2	5.8	0.1	6.1	37.9	10

### **Signalling using VOCs**

The signal emitted by flowers is altered by ozone – altered bVOC quality and quantity. Examples from crops and native (grassland) species

The floral signal is degraded by ozone

More difficult to locate flowers, especially from a distance, which increases foraging times

Foraging time is significantly increased with ozone concentrations of 60 ppb (Fuentes et al., 2016)





### **Direct effects on insects**

Some evidence of direct toxicity at high ozone (100 ppb). At very, very high concentrations ozone can be used as an insecticide for grains.

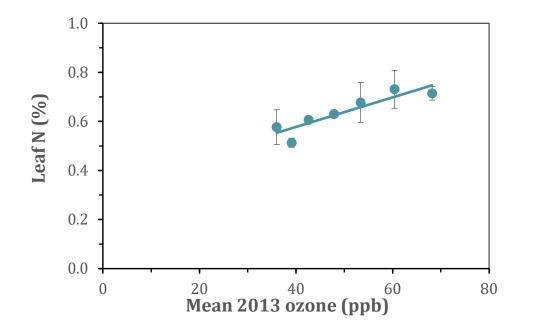
Reduced insect abundance (17%) in Aspen-FACE elevated ozone plots compared to ambient. Generally,  $eO_3$  had the strongest impact on parasitoids (negative) but also influenced some sucking herbivores (positive). (Hillstrom and Lindroth, 2008)

Changes to insect numbers could be mediated by changes in secondary compounds within plant (leaves), e.g. phenols, by changes in the numbers of host plants, or by changes in the habitat quality for predators





### **Ozone impacts on other ecosystem processes**

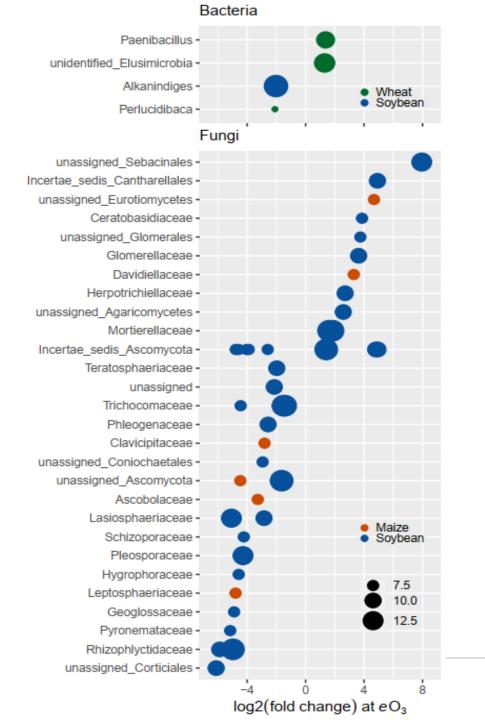


In higher ozone, leaves are lost from trees before most N is re-absorbed.

Higher N content of leaf litter – potential effects on decomposers

Changes in N content and other factors can influence palatability for herbivores. Some (limited) evidence from choice-experiments





## Soil bacteria and fungi

Some evidence of changes to fungal and bacterial diversity in the soil of plant communities exposed to ozone

The role / function of the bacteria/fungi in the ecosystem is not always known.

Sometimes many have similar function

Wang et al., 2022

### **Birds and mammals**



Impacts on birds and mammals are predicted, with mechanisms similar to that for human health.

Bird and mammal populations are declining as a combined result of many stresses.

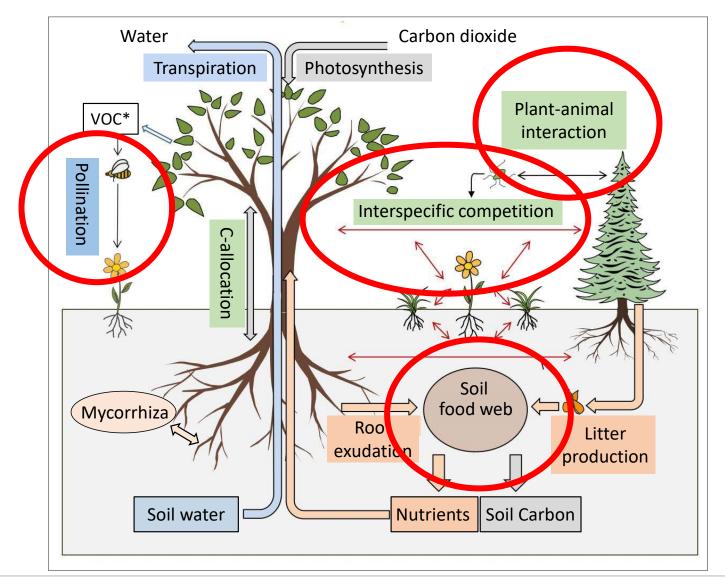
Evidence of impacts is largely from epidemiological studies – and often difficult to disentangle the different pollutants and other stresses.

Ozone has been identified as a contributing factor (Sanderfoot and Holloway, 2017)

Direct effects, but also indirect effects from changes to diet as species at lower trophic levels are impacted, and habitat degradation.



### **Ozone impacts on ecosystem processes**



Modified from Bergmann et al., 2015. http://www.umweltbundesamt.de



### **Summary**

□ There are many effects of ozone identified on plants, which can have direct relevance to biodiversity.

- 'biodiversity' is not always a good metric for impacts (change in species composition)
- □ Effects include ↓ growth including roots, ↓ nitrogen-fixation, altered decomposition/soil cycling, changes in flower number, timing, pollinator signalling.
- □ Effects are difficult and expensive to detect in natural systems
- $\Box$  <1% of plant species have been tested for ozone sensitivity
- □ There is increasing evidence of effects on insects that are mediated via effects on plants
- Evidence on species other than plants and (a few) insects is lacking



# Thank you

Any questions?

